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碩士論文

競爭市場回收清除處理費率和補貼費率的 制定政策

Determining Advanced Recycling Fees and Subsidy Fees in Competitive Closed-loop Supply Chains

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摘要

 近年來,環境保護的概念逐漸形成新的社會認知及規範,是否能妥善地處理經濟快 速發展所造成的電子廢棄物更是廣受注意,因此各國相關的環保法令便因應而生。本研 究利用 Stackelberg 模型求得回收體系參與者的最佳決策。依決策時間點而言,政府(基 管會)為先行者,其先發佈回收清楚處理費率與補貼費率的訊息,跟隨者(責任業者及資 源回收處理業者)接收訊息後,便分別制定最佳製造量及獎勵金水準。其中,為能更貼 近市場現況,我們假設消費市場及回收市場分別存在多家責任業者及資源回收處理業 UITTILI 者,且彼此之間存在著競爭行為。

 許多國家現行的回收制度是採用收支平衡的概念進行回收基金的運用與管理,然 而,政府為非營利的組織,其應以總體社會福利為考量。本研究根據現行收支平衡的概 念以及總體社會福利最大化的目標,於相同稅收水準的假設下求得結果。接著,本研究 利用數值案例進行參數敏感度分析,探討模式可能存在的趨勢。

市場競爭者數量亦會影響到整體社會福利,本研究進一步探討回收基金管理委員會 於回收經營權發放的決策問題,以回收市場中資源回收處理業者的數量對總體社會福利 的影響為探討的主要概念,並佐以數值案例說明模式推演過程。

關鍵字:Stackelberg;回收;競爭模型;費率;回收經營權

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Determining Advanced Recycling Fees and Subsidy Fees in Competitive Closed-loop Supply Chains

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Abstract

The disposal of obsolete electronics products has gained considerable attention due to environmental conservation and legislative requirements. Advanced recycling fees (ARFs) and government subsidy fees may play important roles in recycling. We present a Stackelberg-type model consistings of a leader (the Environmental Protection Administration, EPA) and two followers (MIS firms and recycling firms). MIS firms are the manufacturers, importers, and sellers. The MIS firms and recycling firms both consist of competitive entities. The EPA determines the ARFs paid by the MIS firms and subsidy fees subsidizing recycling firms to maximize the social welfare in closed-loop supply chains, where independent entities maximize their respective profit functions. Then we present a current practice model to determine the fees on the basis of fund balance between revenues and costs. We demonstrate that our results outperform the current practice by a numerical case. We also study how the EPA decides the optimal number of recycling licenses in the recycling market and illustrate the impact of the number of recycling licenses on the value of social welfare, total recycling quantity, reward money, and subsidy fees for the proposed model and the current practice model.

Key words: Advanced recycling fee; Subsidy fee; Stackelberg; Recycle; Closed-loop supply chain; Competitive entities; Recycling license

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李宜庭

誌于 風城交大

中華民國九十九年六月

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Chapter 1 Introduction

The vast consumption of consumer electronics raises the burden on the environment due to the huge amount of obsolete products after usage. The influence of pollution from obsolete electronics products, known as scrap electronics (e-scrap), on the environment is self-evident since they contain metals and other materials that can be hazardous to the environment if they are not properly managed after usage. According to the U.S. Environmental Protection Agency (EPA) study, 40% of the lead in the U.S. landfills is from discarded e-scrap products (DFC, 2009). E-scrap has increased rapidly worldwide. For instance, in developed countries, the average lifetime of a computer is 6-year in 1997 but 2-year in 2005; this change leads to a ballpark number of annual e-scrap generation ranging from 20 to 50 million units (Greenpeace, 2009). In Taiwan, there are about two million e-scrap products recycled according to the Taiwan EPA's statistical data in 2009 (RFMB, 2009a).

In order to relieve the damage to the environment, several regulations are announced. For instance, Waste Electrical and Electronic Equipment (WEEE), Restriction of Hazardous Substances Directive (RoHS), and Eco-Design Requirements for Energy Using Products (EuP) are announced by the European Union. The WEEE indicates that manufacturers bear the responsibility for collecting, recycling, and disposing e-scrap products properly. The RoHS forbids using some specific hazardous substances as raw materials to produce new products. The EuP provides the rules for eco-design to improve the environmental performance of energy-related products (Yen, 2006).

In general, forward supply chains may involve the manufacturing/importing/selling processes of new products, and reverse supply chains may include the reuse/recovery/recycle operations of end-of-life products. In the past decade, much attention has focused on designing proper forward and reverse (closed-loop) supply chains. For example, Fleischmann et al. (2000) derive a classification scheme for different types of recovery networks by comparing the general characteristics of product recovery networks with traditional logistics structures. Guide and Harrison (2003) indicate that new business models need to be developed by cooperating between industry and academia. Wang and Yang (2007) propose a new mixed integer linear programming model to maximize the overall utilization and revenue for designing an e-scrap reverse logistics network. Hong et al. (2006) propose a mixed integer linear programming model to design an infrastructure to process used televisions, monitors, and computer central processing units in the state of Georgia in the U.S to maximize the system net profit, and then robust solution are found with a min–max robust optimization methodology.

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Recycling is a part of the operations in reverse supply chains. It not only decreases the consumption amount of natural resources, but also reduces the impact of obsolete products on the environment. Many researchers have proposed recycling models that maximize total profits and recycling rates by using mathematical programming methodologies (e.g. Inderfurth et al., 2001; Stuart et al.,1999; Uzsoy and Venkatachalam, 1998; Hoshino et al., 1995; Ron and Penev, 1995). Several countries make associated policies to manage the recycling system. For example, the Taiwan EPA imposes taxes, called tax revenues, on the manufacturers, importers and sellers (MIS firms) who are players in forward supply chains. The MIS firms have to pay the e-scrap products processing fee, named as advanced recycling fee (ARF) to support the implement of recycling. On the other hand, consumers may bring the e-scrap products to recycling firms and then receive some reward money paid by the recycling firms. To compensate recycling firms for the costs along with recycling and processing the e-scrap products, the EPA uses the tax revenues to subsidize the recycling firms on the basis of fund balance. The ARF in Taiwan is designed in a similar way to the ARF enacted in California, U.S. The state of California assigns an ARF of \$8-\$25 on all e-scrap products containing hazardous materials depending on the viewable screen size (CalRecycle, 2009). The California EPA uses the tax revenues to establish the Department of toxic substances control (DTSC) besides compensating the recycling firms for the recycling costs incurred. The DTSC is responsible for inspecting the products for hazardous materials (Gable and Shireman, 2001). Canada and Japan have implemented similar programs (Hicks et al., 2005; HP, 2005; Lee et al., 2000; Shih, 2001; Wen, 2005a). However, the EPA is a non-profit organization. It should consider the total social welfare when making policies. It is reasonable to view the EPA as a role of the government, so our model aims to maximize the total social welfare associated with all participants. In general, the social welfare may be defined as the sum of producer surplus, consumer surplus, tax/subsidy revenue, and the environmental externality cost (Bansal and Gangopadhyay, 2003; Hong et al., 2007).

Our modeling framework assumes that the government establishes the associated fees to maximize social welfare and not the fund balance objective in a competitive system. We further assume that the government considers the fees public information and that the associated players select the optimal response to the government-determined rates. Hence, this thesis presents a Stackelberg-type model where the government is a leader to determine the ARFs and subsidy fees, and parties such as MIS firms and recycling firms are the followers, who are competitive participants respectively. The number of recycling firms may affect the value of social welfare, so it is important for the EPA to determine the number of recycling licenses. Therefore, we study how the EPA determines the optimal number of recycling licenses. In this research, we address the following questions:

- (i) Is the concept of fund balance the ideal method for determining the level of ARFs and subsidy fees in a competitive system?
- (ii) What are the socially optimal ARFs and subsidy fees?
- (iii) How might the associated players behave in a competitive system?
- (iv) How might the government behave when it determines the number of WWW. recycling licenses?

The rest of this thesis is organized as follows. In Chapter 2, we review the current environmental policies and the associated instruments. In Chapter 3, we present the social welfare model and fund balance model. Then we solve the optimization problems of these two models for the equilibrium ARFs and subsidy fees established by the EPA and the decisions made by the MIS firms and recycling firms respectively. In Chapter 4, we utilize a case study to examine the difference in the performance measures between the proposed social welfare model and the current practice model. In Chapter 5, we study the impact of the number of recycling licenses on the value of social welfare, total recycling quantity, reward money, and subsidy fees in the recycling market for the social welfare model and the fund balance model. We conclude this thesis in Chapter 6.

Chapter 2 Literature Review

Proper management and recycling of e-scarp products become a challenging issue nowadays. Many countries pay much attention to make associated policies. We review some recycling policies enacted or implemented and the associated instruments in Section 2.1 and 2.2. In this thesis, our model for determining ARF and subsidy fees is to assume that there is a competitive market in closed-loop supply chains. We review some literatures related to the competitive market in Section 2.3.

2.1 The Current Situation of Recycling Policy

In order to effectively regulate the resource recycling activities, there is official regulation on recycling in more and more countries. In 1970, the U.S. EPA was established to encourage reduce/reuse/recycle programs (USEPA, 2009a). The Taiwan EPA established the Recycling Fund Management Board (RFMB) in 1998 for governing the receipt and reimbursement of ARF and subsidy fees, discussing with the recycling firms, subsiding local governments in recycling, and promoting resource recycling activities (RFMB, 2009b).

The implement of extended producer responsibility (EPR) makes manufacturers responsible for the entire lifecycle of the products (Waste to Wealth, 2009). Several countries adopt the concept of EPR and shift the responsibility of recycling on manufacturers. For instance, the Resource Conservation and Recovery Act (RCRA), enacted in 1976, is the principal federal law in the U.S. The RCRA governs the disposal of solid waste and hazardous waste (USEPA, 2009b). Germany issued a regulation called "The closed-loop Economy and Waste Management Act" in 1992. The regulation makes industry responsible for collecting and recycling its products (Fishbein, 1994). In 2002, the Japan Environmental Management Association for Industry (JEMAI) launched "Type III eco-labeling program" which provides quantitative environmental information on products. The program aims to motivate industry to develop, produce and sell eco-friendly products (JEMAI, 2009). In 2006, South Korea tried to carry out "The Act for Resource Recycling of electrical/Electronic Products and Automobiles" which presides over the entire span of product life cycle to promote recycling and restrict the use of hazardous substances from the designing stage of electrical/electronic products and automobiles (IDBMEA, 2009). There are some literatures in support of law enforcement. Foulon et al. (2002) indicate that certain regulation standards need to be imposed to reflect current social responsibility. Chen and Sheu (2009) conclude that governments should gradually raise regulation standards, then the manufacturers gradually improve their product environmental quality, and EPR gets promoted \overline{u} simultaneously.

Other policies involved recycled material flows include taxation and subsidization. Fullerton and Wu (1998) use a simple general equilibrium model to analyze the subsidies for recyclable designs. Conrad (1999) uses a comparative statics analysis to show the impact of a resource and waste taxation on the market volume and the number of firms. Kulshreshtha and Sarangi (2001) show that when consumers bring the reusable part of a product to recycling firms, they must pay a deposit that is subsequently refunded. Bansal and Gangopadhyay (2003) compare different government policies, which are uniform subsidy policy, uniform tax policy, discriminatory tax policy, and discriminatory subsidy policy, and show that a discriminatory subsidy policy is the social welfare improving and also mitigates total pollution. Fullerton and Wolverton (1997) combine environmental subsidy with presumptive tax to propose two-part instrument. The environmental subsidy is provided only to the extent that consumption goods are recycled or that production uses a clean technology, while a presumptive tax is a tax which is imposed under the presumption that all consumption goods become waste or all production uses a dirty technology, which is cheap but, for a given pollution level, it is generally associated with large environmental and enforcement costs (Arguedas, 2005). Two-part instrument internalizes external costs by imposing taxes on the products. Simultaneously, two-part instrument encourages recycling firms to proceed with associated recycling actions through subsidization.

Combining the concept of EPR and two-part instrument, the Taiwan EPA imposes ARFs as recycling funds on MIS firms and then uses the funds to enhance recycling, e.g., subsidizing recycling firms (RFMB, 2009c). On the other hand, recycling firms may compensate the customers, who bring e-scrap products to recycling firms, with a certain amount of reward money to encourage recycling willingness. Similar programs implemented in several other countries can be found in (Hicks et al., 2005; Lee et al., 2000; Wen, 2005a).

2.2 The Instruments of Environmental Policy

Planning and modeling for forward and reverse supply chains have received a growing amount of attention in the past decade. Realff et al. (2004) develop a robust-mixed-integer linear programming model to study a large-scale carpet recycling problem. Nagurney and Toyasaki (2005) construct the multitiered e-cycling network equilibrium model and establish the variational inequality formulation to provide both qualitative properties of the equilibrium pattern as well as numerical examples that are solved using the proposed algorithm. Sheu et al. (2005) consider the used-product return ratio and corresponding subsidies from governmental organizations for reverse logistics to formulate a linear multi-objective programming model. The model optimizes the operations of both integrated logistics and corresponding used-product reverse logistics. They further use a numerical example to indicate that the chain-based aggregate net profits can be improved by 21.1% using the proposed model, compared to the existing operational performance in the particular case studied. Hong et al. (2006) design an infrastructure to process used equipments such as televisions, monitors, and computers in the state of Georgia in the U.S. Wang and Yang (2007) propose a new mixed integer linear programming model to design an e-scrap reverse logistics network. Yang et al. (2009) develop a model of a general closed-loop supply chain network to optimize the equilibrium state of the network by using the theory of variational inequalities. The supply chain network includes raw material suppliers, manufacturers, retailers, consumers and recovery centers. Lee and Dong (2009) propose dynamic location and allocation models to demonstrate the significance of the developed model as well as the efficiency of the proposed solution method. These studies may help the EPA to promote recycling and monitor the whole recycling system.

Several literatures have proposed various environmental instruments in order to reduce the burden of products on the environment. Pigou (1920) proposes the concept of economic externalities and Pigovian tax. A Pigovian tax is a tax imposed on a non-market activity which causes negative externalities. Jung et al. (1996) internalizes environmental pollution costs by imposing taxes on enterprises to study

the incentives for advanced pollution abatement technology at the industry level. In 2008, the Canadian province of British Columbia announced that it was going to impose a carbon tax, which is used for achieving the environment conservation by reducing the emissions of carbon dioxide, of \$10 on per metric ton of carbon dioxide equivalent emissions (Canada, 2009). However, it is difficult to estimate and control the environmental externality cost. Cremer and Thisse (1999) indicate that a country may spend much money on estimating environmental externality cost. Hence, the concept of a Pigovian tax is difficult to be implemented in practice. On the other hand, the required data of imposing presumptive taxes on income are price and demand. Because the required data of imposing presumptive taxes are more available and correct than the required data of imposing Pigovian taxes, the concept of presumptive tax is adopted in some associated literatures. Bansal and Gangopadhyay (2003) study policy measures to improve environmental quality. Besides, Bansal and Gangopadhyay (2003) also focus on the manufacturers, who do not place much importance on environmental conservation, and then impose taxes on the products according to their production quantity.

Many economists consider that the concept of subsidization should be added into the environmental instruments because the economic incentives are not powerful influences on recycling while the government imposes taxes only on the MIS firms. For example, Fullerton and Wolverton (1997) build two simple general models to demonstrate that two-part instrument is easier to implement compared to the Pigovian tax. Bansal and Gangopadhyay (2003) find that a uniform subsidy policy improves the average environmental quality while a uniform tax policy worsens it, and a discriminatory subsidy policy reduces total pollution and enhances aggregate welfare while a discriminatory tax policy may increase total pollution and reduce aggregate welfare. Wen (2005b) proposes that the recycling firms are willing to invest in recycling equipment with a discriminatory subsidy policy. Furthermore, there are some literatures which combine taxation with subsidization to propose different objective functions. Wen (2005a) integrates the model of Fullerton and Wolverton (1997) into the recycling system and explains that the social welfare can be achieved if the government imposes taxes on the MIS firms while subsidizing recycling firms for the costs caused by processing the obsolete products.

2.3 The Competitive Market

Researches on recycling policy are mainly based on a single company model where each participant, such as the MIS firms or recycling firm is a single entity (Fullerton and Wu, 1998; Choe and Fraser, 2001; Stavins, 2002). However, there are different numbers of firms competing in the real-world. In a competitive market, independent entities maximize their own profit functions respectively and are unwilling to reveal 47111111 private information to others.

The assumptions of emerging literatures on reverse supply chains are mostly assumed to be competitive. Jung et al. (1996) evaluate the incentive effects of five environmental policy instruments, which are performance standards, emissions subsidies, emissions taxes, and issued and auctioned marketable permits, to promote the development and adoption of advanced pollution abatement technology in a heterogeneous and competitive industry. Kfiberger and Karlsson (1998) propose that the data, which is from specific, contracted electricity production plants, should be used for electricity consumption in lifecycle analysis. And the electricity is purchased from a competitive market. Majumder and Groenevelt (2001) present a two-period model to examine the effect of competition in remanufacturing. Mitra and Webster (2008) also analyze a two-period model of a manufacturer and a remanufacturer and show that the introduction of partial subsidies increases both manufacturer's and remanufacturer's profits. Chen and Sheu (2009) design proper environmental-regulation pricing strategies for green supply chain firms in a competitive market.

A literature survey reveals little research on how ARF and subsidy fees are determined in a competitive market where there are several individual firms in each tear. Hong et al. (2007) view the MIS firms and recycling firms as two separate parties. In the real-world, there are several firms competing in the tiers of MIS firms and recycling firms. In this research, we study how the EPA establishes the best associated fees when there are many MIS firms and recycling firms.

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For some industries, the enterprises have to be granted legal licenses by the government. The competitiveness between enterprises may be induced by the number of licenses (Jehiel and Moldovanu, 2000). Jehiel and Moldovanu (2000) analyze the interplay between the number of 3G licenses and the market structure in a model with several incumbents and several potential entrants. They show that plausible conditions under which all incumbents get a license, and more licenses need not result in greater competitiveness if the number of incumbents is greater than the number of new licenses. In e-scrap reverse supply chains, the EPA uses the ARFs paid by the MIS firms to compensate recycling firms for the operational and recycling costs incurred, and the recycling firms have to possess legal recycling licenses to obtain subsidies from the government for costly recycling operations. For example, the Taiwan recycling firms are required to be granted recycling licenses and then subsidized by the EPA (RFMB, 2009d). In this thesis, we study how the EPA determines the optimal number of recycling licenses in a recycling market to maximize the social welfare

Chapter 3 The Model

We present a Stackelberg-type model to describe a competitive closed-loop (forward and reverse) supply chain system consisting of the government (EPA), manufacturers, importers, or sellers (MIS firms), and recycling firms (rec). The MIS represents the associated entities involved in forward supply chains, and the recycling firms include collection, consolidation, or processing sites in reverse supply chains. The EPA determines ARFs and subsidy fees to maximize social welfare, and the MIS firms and recycling firms seek their own objectives which respond to the EPA-determined rates. It is reasonable to assume that the EPA acts as leader, and the MIS firms and recycling firms are two followers. In this thesis, we refer to the model proposed in Section 3.2 as the social welfare model.

According to the current practice, the EPA determines ARFs and subsidy fees on the basis of fund balance between revenues and costs along with recycling operations. For comparative purposes, we construct a fund balance model, where the total revenue the EPA collects equals the EPA's total expenditure, to use as a benchmark to compare with the social welfare model. The fund balance model is described in Section 3.3.

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The major difference between these two models is the objective function where the social welfare model aims to maximize the social welfare and the fund balance model determines the ARFs and subsidy fees on the basis of fund balance between tax revenues and subsidy expenditures along with recycling operations.

3.1 Supply Chain Flows

There are three key elements describing our supply chain system: material, cash, and information flows. In this thesis, we assume that a supply chain consists of three groups: MIS firms, customers, and recycling firms. In general, the MIS firms may act as manufacturers, importers, or sellers selling electronics products to customers. After usage, customers may bring obsolete products to recycling firms which remanufacture or recycle the e-scrap products and convert them into recovery materials as well as some accompanying trash. We reasonably assume that the MIS firms exist in one market. For example, there are different MIS firms in Taiwan market for laptop computers; that is, the MIS firms exist in a competing market. On the other hand, we reasonably assume that the recycling firms exist in distinct market segments or distinct geographic locations (Hong et al., 2008). About the cash flow, when the MIS firms manufacture, import, or sell electronics products, they pay the ARFs according to electronics production quantity, in support of the implementation of e-scrap recycling. On the other hand, the EPA uses the ARFs to subsidize recycling firms according to recycling quantity for the operational and recycling costs incurred. Then recycling firms may compensate customers with a certain amount of reward money to encourage recycling behavior. There are two stages describing the information flow according to the timeline. The first stage is that the EPA announces the ARFs and subsidy fees to the public. The second stage is that the MIS firms and recycling firms determine their own optimal policies after observing the rates announced by the EPA. In particular, we assume that the MIS firms and recycling firms both consist of independent and competitive entities respectively. Each entity maximizes its own profit function and is unwilling to reveal its private information to others. The flows of these three elements are represented in Figure 1 where there are

3.2 The Social Welfare Model

The EPA is a unit of the government and it is reasonable to assume that the EPA should consider the social welfare when it makes policies. We present a Stackelberg-type model in a competitive reverse supply chain where each participant independently acts according to its own interests. The leader considers the followers' potential decisions by anticipating followers' behavior, and then makes its optimal policy. The followers may make their optimal policies according to the policy announced by the leader.

In addition, we note there are two markets in the proposed model: one is the

consuming market, where new products are sold to customers, and the other one is the recycling market, where obsolete products are brought to recycling firms for recycling. We assume that there are competitive participants in the consuming market and in the recycling market. Furthermore, each participant aims to maximize its profit. It is reasonable to assume that the number of participants and the profit function of each participant are common knowledge - a typical assumption in the game-theoretically type model (Gibbons, 1992). Our model can be classified as a two-stage dynamic game of complete information. A common method for solving this problem is the backward induction technique, which is the process of reasoning backwards in time, from the end of a problem, to determine a sequence of optimal actions (Von Neumann and Morgenstern, 1994). In this study, we apply the backward induction to solve the proposed model.

3.2.1 The Model of MIS Firms

We first construct the MIS model to determine the electronics production quantity given the ARFs announced by the EPA. We assume the market, where the electronics products are manufactured, imported, and sold, consists of *n* MIS firms and each one aims to maximize its own profit. In addition, the MIS firms make their decisions simultaneously.

Let q_{x_i} denote the *i* th MIS firm's production quantity, $i = 1, 2, 3, ..., n$. The total demand in the market, Q_x , are the sum of MIS firms' production quantity, that is, $\frac{1}{1}$ $\frac{1}{x_i}$ *n* $x = \angle 4x$ *i* $Q_x = \sum q$ $=\sum_{i=1}^{n} q_{x_i}$. Assume that the total demand is characterized by a commonly-used linear demand function, $P_x = a - bQ_x$, where P_x is the market price, *a* is the

intercept parameter, and *b* is the slope parameter, $a, b > 0$. It means that when one unit of products is produced, the market price is decreasing in *b* units of market price. In other words, if *b* is a large number, the market price may decrease rapidly with increasing production quantity. A linear form of the inverse demand function helps us obtain qualitative insights without much analytical complexity.

The MIS firms' production processes and skills are different, so the unit production cost of each MIS firm is reasonable assuming non-identical. Let C_v be the *i* th MIS firm's unit production cost. In addition, the MIS firms pay the ARF, denoted by *t* per unit of products, in support of the implementation of the recycling program. The profit function of the *i* th MIS firm, denoted by \prod_{MIS_i} , is

⁰ Max () . ¹ *i ii xi MIS x v x ^q P C tq* (3.1)

It is reasonable to assume that the number of MIS firms and each MIS firm's profit function are common knowledge among all MIS firms. Equation (3.1) may be transformed into a one-variable function. Substituting the demand function, $P_x = a - bQ_x$, in (3.1) results in

$$
\max_{q_{x_i} \geq 0} \ \prod_{MIS_i} = (a - b \sum_{i=1}^n q_{x_i} - C_{v_i} - t) q_{x_i}.
$$
 (3.2)

The profit function (3.2) is concave in q_x , whenever $b > 0$, so (3.2) is maximized when the first-order condition holds, i.e. when

$$
q_{x_i}^* = \frac{1}{2b} \left(a - t - C_{y_i} - b \sum_{\substack{j=1 \ j \neq i}}^n q_{x_j}^* \right).
$$
 (3.3)

Equation (3.3) specifies each MIS firm's best response to the information announced by the EPA, i.e. the level of the ARFs, *t* . In game theory, the Nash equilibrium is a solution concept of a game involving two or more players, in which no player has incentive to deviate from his/her action given that the other players do not deviate.

In other words, in the Nash equilibrium solution, no one can be better off by a unilateral change in its solution (Gibbons, 1992). According to the concept of the Nash equilibrium, we combine these n best response equations to n -variable simultaneous equations to solve for the equilibrium solution. First, we add up these *n* equations, i.e. (3.3), together and solve for the total demand as follows:

$$
Q_x^* = \frac{1}{2b} \left(n(a-t) - \sum_{i=1}^n C_{\nu_i} - (n-1)bQ_x^* \right).
$$
 (3.4)

Rewriting (3.4), we have

$$
Q_{x}^{*} = \frac{1}{(n+1)b} \left(n(a-t) - \sum_{i=1}^{n} C_{\nu_{i}} \right).
$$
 (3.5)

Substituting $Q_x^* = \sum q_x^*$ $\frac{1}{1}$ ^{1 *x*_i} *n* $x - y$ $\frac{1}{4}$ *i* $Q_x^* = \sum q$ $=\sum_{i=1}^{n} q_{x_i}^*$ in (3.3), we have * $\frac{d}{dx} = \frac{a + b + b}{b}$. $v_i - \nu \mathcal{Q}_x$ *x* $a - t - C_{v_i} - bQ$ $q_{x_i}^* = \frac{a + b}{b}$ $=\frac{a-t-C_{v_i}-bQ_x^*}{\sqrt{2\pi}}$ (3.6)

Substituting (3.5) in (3.6), we obtain each MIS firm's best response to t .

$$
q_{x_i}^* = \frac{1}{(n+1)b} \left(a - t - nC_{v_i} + \sum_{\substack{j=1 \ j \neq i}}^n C_{v_j} \right) \qquad (3.7)
$$

Equation (3.7) specifies the *i* th MIS firm's optimal production quantity after it observes the level of the ARF, *t* , announced by the EPA. In other words, the MIS firm's production quantity, $q_{x_i}^*$, is a function of the ARF rate, *t*, and (3.7) can be rewritten as $q_{x_i}^*(t)$. In addition, substituting (3.5) in the demand function results in the market price.

$$
P_x^* = \frac{1}{(n+1)} \left(a + nt + \sum_{i=1}^n C_{v_i} \right)
$$
 (3.8)

3.2.2 The Model of Recycling Firms

Now consider the model of recycling firms to determine the rate of reward money given the level of subsidy fee announced by the EPA. Recycling firms may compensate customers, who bring e-scarp products to recycling firms, with a certain amount of reward money to encourage recycling after usage. A reasonable customer is more willing to bring e-scarp products to the recycling firms with higher rewards, so we assume that the relationship between the recycling quantity and reward money is positive, linear in a competitive market where there are *m* recycling firms. From (Hong et al., 2008), we understand that recycling firms usually collect e-scrap products in distinct market segments or distinct geographic locations. Moreover, different market areas cannot be simplified as one single market, so we let c_i and d_j denote the *j* th recycling firm's market intercept parameter and slope parameter respectively, $j = 1, 2, \dots, m$. Intuitively, high reward money may increase customers' willingness to bring e-scrap products to recycling firms.

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The reward money determined by other recycling firms may affect the *j* th recycling firm's recycling quantity. In addition, the customers' recycling behavior is influenced by not only the amount of reward money but also other factors such as the residential regions. For example, a customer may not bring e-scrap products to the recycling firm, whose location is far away from the customer's location, even if its reward money is high. Therefore, we let k_i^j denote the decrease in the recycling quantity in the *j* th recycling firm caused by a unit of increase in the reward money paid by the *l* th recycling firm, $k_i^j \ge 0$. Let P_{w_i} and q_{c_i} denote the *j* th recycling firm's reward money paid to customers and recycling quantity respectively.

The relationship is listed as follows:

$$
q_{c_j} = c_j + d_j P_{w_j} - \sum_{\substack{l=1 \ l \neq j}}^m k_l^j P_{w_l}.
$$
 (3.9)

We refer to (3.9) as the recycling quantity function. Similar modeling ways to (3.9) can be found in (Gibbons, 1992; Toyasaki et al., 2008; Majumder and Groenevelt, 2001).

Recycling firms have respective skills and processes, so the unit processing cost of recycling firms is not identical. Let r_i denote the *j* th recycling firm's net cost for recycling one unit of e-scrap products, $r_j > 0$. On the other hand, recycling firms' revenues are the subsidy fees subsidized by the EPA. Let *s* denote the subsidy fee per unit of e-scrap products. The profit function of the *j* th recycling firm is

$$
\mathbf{Max}_{P_{w_j} \geq 0} \quad \Pi_{rec_j} = (s - r_j - P_{w_j})q_{c_j}.
$$
 (3.10)

It is reasonable to assume that the number of recycling firms and each recycling firm's profit function are common knowledge among all recycling firms. Like the MIS firms, recycling firms simultaneously make their own decisions. Substituting (3.9) in (3.10), the profit function is concave in P_{w_j} , whenever $d_j > 0$. Equation (3.10) is maximized when the first-order condition holds, i.e. when

$$
P_{w_j}^* = \frac{1}{2d_j} \left(d_j (s - r_j) - c_j + \sum_{\substack{l=1 \ l \neq j}}^m k_l^j P_{w_l}^* \right). \tag{3.11}
$$

Let the *y* th recycling firm represents one of the recycling firms, and it is not the *j* th recycling firm, $y = 1, \dots, j-1, j+1, \dots, m$. From (3.11), the *y* th recycling firm's best response to *s* is

$$
P_{w_y}^* = \frac{1}{2d_y} \left(d_y (s - r_y) - c_y + \sum_{\substack{l=1 \ l \neq y}}^m k_l^y P_{w_l}^* \right). \tag{3.12}
$$

Subtracting (3.12) from (3.11) , we have

$$
P_{w_j}^* = \frac{1}{(2d_j + k_j^y)} \left(\frac{(2d_y + k_y^j) P_{w_y}^* + (d_j(s - r_j) - c_j) - (2d_j + k_j^y)}{(d_y(s - r_y) - c_y) - \sum_{\substack{l=1 \ l \neq j, y}}^m P_{w_l}^*(k_l^y - k_l^j)} \right). \tag{3.13}
$$

From (3.13), we obtain the relationship between the *j* th and *y* th recycling firms' decisions where the *y* th recycling firm can be viewed as one of any particular recycling firm other than the *j* th recycling firm. If we directly use (3.13) to solve for the reward money response to *s* , we cannot find a closed-form solution of reward money. In order to simplify this model, we assume that the parameter k_i^j is set according to the market situation. More specifically, k_i^j is set according to the size of the market area where the *l* th recycling firm exists in the recycling market. Therefore, the parameter k_i^j is a fixed value while *l* is fixed and *j* is one of the other recycling firms; that is, $j \in \{1, 2, \dots, l-1, l+1, \dots, m\}$. From this condition follows Assumption 1.

Assumption 1 *We let the value of* k_i^j *be the same, where l is fixed and j is not fixed,* $j \in \{1, 2, \dots, l-1, l+1, \dots, m\}$. That is,

$$
k_l^1 = k_l^2 = \dots = k_l^m. \tag{3.14}
$$

Substituting (3.14) in (3.13),

$$
P_{w_y}^* = \frac{(2d_j + k_j^y)P_{w_j}^* - (d_j(s - r_j) - c_j) + (d_y(s - r_y) - c_y)}{2d_y + k_y^j}.
$$
\n(3.15)

Substituting (3.15) in (3.11) and denoting $A = 2d_i - \sum_{i=1}^{m} \frac{k_i^j (2d_j + k_j^l)}{2d_j + k_j^l}$ 1 $2d_i +$ 2 $2d_1 +$ $\sum_{l}^{m} k_l^{j} (2d_j + k_j^l)$ $j = \sum_{\substack{l=1 \ l \neq j}}$ $2d_l + k_l^{j}$ $k_l^j(2d_j+k)$ $A = 2d$ $\sum_{i=1}^{l} 2d_i + k$ $=2d_j-\sum_{j} \frac{\kappa_i(2a_j+\kappa_j)}{2d+k^j}$ in order to

simplify our notation, we obtain the *j* th recycling firm's best response to *s* .

$$
P_{w_j}^* = \frac{1}{A} \left(\sum_{\substack{l=1 \ l \neq j}}^M \frac{k_l^j \left(\left(d_j \left(s - r_j \right) - c_j \right) - \left(d_l \left(s - r_l \right) - c_l \right) \right)}{2 d_l + k_l^j} \right) \tag{3.16}
$$

Equation (3.16) specifies the *j* th recycling firm's optimal reward money after it observes the level of the subsidy fee, *s* , announced by the EPA. In other words, the recycling firm's reward money, $P_{w_i}^*$, is a function of the subsidy fee rate, *s*, and (3.16) can be rewritten as $P_{w_i}^*(s)$. Rewriting (3.11), we have

$$
\sum_{\substack{l=1 \ l \neq i}}^m k_l^j P_{w_l}^* = 2d_j P_{w_j}^* - d_j (s - r_j) + c_j.
$$
\n(3.17)

Substituting (3.17) in (3.9) , we have

$$
q_{c_j}^* = d_j (s - r_j - P_{w_j}^*).
$$
 (3.18)

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Substituting (3.16) in (3.18) , we obtain the *j* th recycling firm's resulting recycling quantity.

$$
q_{c_j}^* = d_j (s - r_j) - \frac{d_j}{A} \left(\sum_{\substack{l=1 \ l \neq j}}^m \frac{k_l^j \left(\left(d_j (s - r_j) - c_j \right) - \left(d_l (s - r_l) - c_l \right) \right)}{2d_l + k_l^j} \right) \tag{3.19}
$$

3.2.3 The Model of the EPA

The main objective of the EPA is to maximize the social welfare, which is the sum of the producer surplus, consumer surplus, tax/subsidy revenue, and the environmental externality cost (Bansal and Gangopadhyay, 2003). The producer surplus here is the sum of the profits of the MIS firms, $\sum (P_{x} - C_{y_{i}} - t)$ $\mathbf{v}_i \cdot \mathbf{x} = \mathbf{v}_i \cdot \mathbf{y}$ *n* $\mathbf{v}_x - \mathbf{v}_{v_i} - \mathbf{v}_j \mathbf{q}_x$ *i* $P_x - C_{v_i} - t$ *q* $\sum_{i=1} (P_{x} - C_{v_i} - t) q_{x_i}$, and the profits of recycling firms, $\sum (s - P_{w_i} - r_i)$ $\mathbf{u}_j \begin{pmatrix} z & -w_j & y \end{pmatrix} \mathbf{u}_j$ *m* $w_i = I_j$ μ_c *j* $s - P_{w_i} - r_i$ $\sum_{j=1} (s - P_{w_j} - r_j) q_{c_j}$. Hence, the producer surplus is

$$
\sum_{i=1}^{n} \left(P_x - C_{\nu_i} - t \right) q_{x_i} + \sum_{j=1}^{m} \left(s - P_{\nu_j} - r_j \right) q_{c_j}.
$$
 (3.20)

The consumer surplus is the sum of the consumer surplus in the consuming market and in the recycling market. In a consuming market, the consumer surplus is the difference between the price that consumers are willing to pay and the actual market price. In other words, the consumer surplus is the triangular area above the market price level and below the demand curve. In a recycling market, if the announced rate of reward money is greater than the fee level that consumers are willing to be paid to bring their obsolete products to recycling firms. The difference between the reward money that consumers are willing to be paid and the actual reward money is the consumer surplus in recycling market. From Figure 2, we obtain the consumer surpluses. Let *CS*₁ represents the consumer surplus in the consuming market and $CS₂$ be the consumer surplus in the recycling market, we have

$$
CS_1 = \frac{1}{2}bQ_x^2,\tag{3.21}
$$

$$
CS_2 = \sum_{j=1}^{m} \left(P_{w_j} q_{c_j} - \frac{1}{2} d_j P_{w_j}^2 \right). \tag{3.22}
$$

Figure 2 The consumer surplus in the consuming market and in the recycling market

The EPA imposes ARFs on the MIS firms based on the production quantity. The total ARFs are called total tax revenue (TTR), and its value is $\frac{1}{1}$ ^{$\frac{1}{x}$} *n x i* $t \sum q$ $\sum_{i=1} q_{x_i}$. As mentioned earlier, the EPA uses the ARFs to compensate recycling firms for associated recycling costs. The total expenditure used to subsidize recycling firms

are called total subsidy expenditure (TSE), and its value is $\frac{1}{1}$ ^{1c}_j *m c j s q* $\sum_{j=1}^{n} q_{c_j}$. The tax/subsidy

revenue is the total tax revenue in the consuming market minus the total subsidy expenditure in the recycling market.

$$
t\sum_{i=1}^{n}q_{x_i}-s\sum_{j=1}^{m}q_{c_j}
$$
\n(3.23)

The environmental externality cost is a detrimental impact on a party and not directly involved in an economic transaction (Koomey and Krause, 1997). There are two types of environmental externality cost in our model. One is the indirect pollution costs resulting from producing new products. Let *e* denote the unit indirect pollution cost incurred in producing new products, so the pollution cost resulting from producing new products is eQ_x . The other one is the pollution cost caused by uncollected e-scrap products. Let *E* denote the unit pollution cost of uncollected e-scrap products. Before evaluating the pollution costs of uncollected e-scrap products, we should estimate the total amount of e-scrap products of current generation. In practice, the amount of e-scrap products may not be available to decision-makers, or it is difficult to estimate. Instead, the amount of current generation of new electronic products is relatively traceable and probably can be obtained from the associated government agencies, such as the department of commerce. We characterize the return rate of e-scrap products by τ , $0 < \tau < 1$, the

rate of current generation of new products that are expected to return to the reverse channel after usage, so the total e-scrap products of current generation are τQ_x , and

the pollution costs caused by uncollected obsolete products are $\frac{1}{1}$ ^{1c}_j *m* $x = \angle q_c$ *j* $E| \tau Q_x - \sum q$ $\bigg(\tau Q_{\scriptscriptstyle \cal X} - \sum_{\scriptscriptstyle j=1}^m q_{\scriptscriptstyle \cal C_j}\,\bigg).$

The models with the similar concept of return rate appear in (Savaskan et al., 2004; Savaskan and Van Wassenhove, 2006). Adding up these two types of environmental externality cost together and the total environmental externality cost can be described as

$$
E\left(\tau Q_{x}-\sum_{j=1}^{m}q_{c_{j}}\right)+eQ_{x}.
$$
\n(3.24)

From (3.20), (3.21), (3.22), ((3.23), and (3.24), the EPA optimizes the total social welfare as shown in (3.25).

$$
\begin{split} \max_{t,s\geq 0} \ \prod_{\substack{g_{ov}}} \ \sum_{i=1}^{n} \Big(P_x - C_{v_i} - t \Big) q_{x_i} + \sum_{j=1}^{m} \Big(s - P_{w_j} - r_j \Big) q_{c_j} + \frac{1}{2} b Q_x^2 + \\ \sum_{j=1}^{m} \Big(P_{w_j} q_{c_j} - \frac{1}{2} d_j P_{w_j}^2 \Big) + \Bigg(t \sum_{i=1}^{n} q_{x_i} - s \sum_{j=1}^{m} q_{c_j} \Bigg) - E \Bigg(\tau Q_x - \sum_{j=1}^{m} q_{c_j} \Bigg) - e Q_x \end{split} \tag{3.25}
$$

With that rationality assumption, the EPA anticipates that the MIS firms and recycling firms choose their optimal response to the announced fees. This allows us to substitute (3.7) , (3.8) , (3.16) , and (3.19) in (3.25) . Again to simplify the notation in

this thesis, we denote
$$
B = d_j - \sum_{\substack{l=1 \ l \neq j}}^m \frac{k_l^j (d_j - d_l)}{2d_l + k_l^j}
$$
 and

$$
C = d_j r_j + c_j + \sum_{\substack{l=1 \ l \neq j}}^m \frac{k_l^j \left(-d_j r_j - c_j + d_l r_l + c_l \right)}{2d_l + k_l^j}.
$$
 Equation (3.25) is maximized when

the first-order condition holds, i.e. when

$$
t^* = \frac{1}{n^2} \bigg((n+1)(nE\tau + ne) - na + \sum_{i=1}^n C_{v_i} \bigg), \tag{3.26}
$$

$$
s^* = \frac{\sum_{j=1}^m \left(d_j \left(E - r_j + \left(\frac{B}{A} \right) \left(r_j - E + \frac{C}{A} \right) \right) \right)}{\sum_{j=1}^m \left(d_j \left(\frac{B}{A} \right)^2 \right)}.
$$
\n(3.27)

The profit function, $\prod_{g \circ v}$, is concave both in *t* and *s*, whenever $b > 0$, $d_i > 0$.

The second-order conditions result in

$$
\frac{\partial^2 \Pi_{gov}}{\partial t^2} = -\frac{n^2}{b(n+1)^2},\tag{3.28}
$$

$$
\frac{\partial^2 \Pi_{\text{gov}}}{\partial s^2} = -\sum_{j=1}^{m} \left(d_j \left(\frac{B}{A} \right)^2 \right).
$$
\n(3.29)

That is, the values of (3.28) and (3.29) are both negative.

In addition, substituting (3.26) in (3.7) and (3.8), we obtain the *i* th MIS firm's optimal production quantity and market price, $i = 1, 2, 3, ..., n$.

∥∃

$$
q_{x_i}^* = \frac{1}{n^2 (n+1)b} \left(\sum_{i=1}^n C_{v_i} - n^3 C_{v_i} + n^2 \sum_{\substack{j=1 \ j \neq i}}^n C_{v_j} \right) \tag{3.30}
$$

$$
P_{x}^{*} = E\tau + e + \frac{1}{n} \sum_{i=1}^{n} C_{\nu_{i}}
$$
 (3.31)

Substituting (3.27) in (3.16) and (3.19), we obtain the *j* th recycling firm's optimal reward money and resulting recycling quantity, $j = 1, 2, \dots, m$.

$$
P_{w_j}^* = \frac{1}{A} \left[\frac{d_j \sum_{j=1}^m d_j (E - r_j + \left(\frac{B}{A}\right) (r_j - E + \frac{C}{A}))}{\sum_{j=1}^m d_j \left(\frac{B}{A}\right)^2} - d_j r_j - c_j - \frac{1}{A} \right]
$$
\n
$$
P_{w_j}^* = \frac{1}{A} \left[\frac{1}{\sum_{l=1}^m \frac{k_l^j}{2d_l + k_l^j}} \left[(d_j - d_l) \frac{\sum_{j=1}^m d_j (E - r_j + \left(\frac{B}{A}\right) (r_j - E + \frac{C}{A}))}{\sum_{j=1}^m d_j \left(\frac{B}{A}\right)^2} \right] \right] \tag{3.32}
$$
\n
$$
q_{c_j}^* = -d_j r_j + d_j \left(1 - \frac{1}{A} \right) \left(\frac{d_j \sum_{j=1}^m d_j (E - r_j + \left(\frac{B}{A}\right) (r_j - E + \frac{C}{A}))}{\sum_{j=1}^m d_j \left(\frac{B}{A}\right)^2} \right) + \frac{1}{A} \left[\frac{d_j r_j + c_j + \frac{1}{A} \left[(d_j - d_l) \frac{\sum_{j=1}^m d_j (E - r_j + \left(\frac{B}{A}\right) (r_j - E + \frac{C}{A}))}{\sum_{j=1}^m d_j \left(\frac{B}{A}\right)^2} - \frac{1}{A} \right] \right] \tag{3.33}
$$
\n
$$
d_j \left(\frac{m}{\sum_{l=1}^m \frac{k_l^j}{2d_l + k_l^j}} \left((d_j - d_l) \frac{\sum_{j=1}^m d_j (E - r_j + \left(\frac{B}{A}\right) (r_j - E + \frac{C}{A}))}{\sum_{j=1}^m d_j \left(\frac{B}{A}\right)^2} - \frac{1}{A} \right) \right]
$$
\n
$$
(3.33)
$$

The subgame perfect Nash equilibrium (SPNE) is equilibrium such that players' strategies constitute a Nash equilibrium in every subgame of the original game. The SPNE is normally deduced by the backward induction (Gibbons, 1992). To calculate the SPNE, the best response functions of the followers, i.e. the MIS firms and recycling firms, must first be calculated given the output of the leader, i.e. the EPA. According to Section 3.2.1 and Section 3.2.2, the best response functions of the MIS firms and recycling firms are represented by $q_{x_i}^*(t)$ and $P_{w_i}^*(s)$, respectively. Then the EPA's equilibrium outcomes that maximize the social welfare are solved given $q_{x_i}^*(t)$ and $P_{w_i}^*(s)$, i.e. given the best responses of the MIS firms and recycling firms. According to Section 3.2.3, the equilibrium outcomes of the EPA are t^* and s^* . Therefore, the solution $(t^*, s^*, q^*_{x_i}(t), P^*_{w_i}(s))$ represented in equations (3.26), (3.27),

(3.30), (3.32) is the SPNE of the social welfare model. Under the SPNE, no one has anything to gain by changing only his or her own strategy unilaterally. Hence, we obtain the two fees which achieve the maximum of the social welfare at the equilibrium status, while both the MIS firms and recycling firms gain the maximum of their profits under their best responses to any possible pair of the ARF and subsidy fee (t, s) .

3.3 The Fund Balance Model

The current practice in the e-scrap recycling systems in Taiwan and the state of California in the U.S. determines the ARF and subsidy fee on the basis of fund balance between revenues and costs along with recycling operations (Lee et al., 2000; IWMB, 2003). In this section, we develop a fund balance model, where the total revenue the EPA collects equals the EPA's total expenditure, as a benchmark to compare with the social welfare model in Section 3.4. The material, cash, and information flows of the fund balance model are the same as the supply chain flows in the social welfare model and represented in Figure 1. In addition, we let the EPA collect the identical total tax revenue in these two models and let all parameters in the fund balance model be the same as the parameters in the social welfare model in order to have a fair basis for comparison.

3.3.1 The Model of MIS

Let t' denote the ARF per unit paid by the MIS firms in the fund balance model. The decision variable in this section is also the production quantity. Let q_{x_i}

denote the *i* th MIS firm's optimal production quantity and $\Gamma = \sum_{i=1}^{\ } q_{x_{i}} \Gamma_{i}^{x_{i}}$ *n* α *z* α *z* α *i* $Q_x = \sum q$ $=\sum_{i=1}^{n} q_{x_i}$ ' is the total demand for new products, which is characterized by a commonly-used linear demand function, $P_x' = a - bQ_x'$, where P_x' is the market price. Parameters, *a* and *b*, are the same as the parameters in the social welfare model. The *i* th MIS firm's profit function framework is not changed, $i = 1, 2, \dots, n$.

$$
\mathbf{Max}_{q_{x_i} \geq 0} \quad \prod_{MIS_i} = (P_x - C_{v_i} - t')q_{x_i}
$$
\n(3.34)

Using similar computational procedure in the social welfare model, we obtain the *i* th MIS firm's optimal production quantity and market price as

$$
P_{x} = \frac{1}{(n+1)} \left(a + nt + \sum_{i=1}^{n} C_{v_i} \right),
$$
\n(3.35)

$$
q_{x_i}^{*} = \frac{1}{(n+1)b} \left(a - t - nC_{y_i} + \sum_{j=1}^n C_{y_j} \right). \tag{3.36}
$$

Adding up MIS firms' production quantity together, we obtain the total demand for $\frac{1}{2}$ new products as follows:

$$
Q_x^{*} = \frac{1}{(n+1)b} \left(n(a-t') - \sum_{i=1}^{n} C_{y_i} \right)
$$
 (3.37)

3.3.2 The Model of Recycling Firms

In the current recycling policy, the government's subsidy funds are not all received by the recycling firms. Part of the subsidy funds may be received by the local waste management under the local government supervision and part may be received indirectly by communities to proceed with recycling. Because the functional characteristics of recycling organizations are similar, we conceptually view those recycling organizations as individually-owned recycling firms in the group of recycler. Let *s*' denote the subsidy fee per unit, P_{w_j} and q_{c_j} denote the *j* th recycling

firms optimal reward money and resulting recycling quantity, and $\Gamma = \sum_{j=1}^r {q}_{c_j} \; \Gamma$ *m* $c = \sum q_c$ *j* Q_c = $\sum q$ $= \sum_{j=1}^{ } q_{c_j}$ ' be the total recycling quantity. The parameters c_j , d_j , r_j , and k_i^j are the same as the parameters in the social welfare model. The j th recycling firm's profit function framework is not changed, $j = 1, 2, \dots, m$.

$$
\underset{P_{w_j} \geq 0}{\text{Max}} \ \Pi_{\text{rec}_j} = (s - r_j - P_{w_j}) q_{c_j} \tag{3.38}
$$

Using similar computational procedure in the social welfare model, we obtain the *j* th recycling firm's optimal reward money and resulting recycling quantity.

$$
P_{w_j}^{*} = \frac{1}{A} \left(\sum_{\substack{l=1 \ l \neq j}}^{m} \frac{k_l^j \left(\left(d_j \left(s - r_j \right) - c_j \right) - \left(d_l \left(s - r_l \right) - c_l \right) \right)}{2d_l + k_l^j} \right), \tag{3.39}
$$
\n
$$
q_{c_j}^* = d_j \left(s - r_j \right) - \frac{d_j}{A} \left(\sum_{\substack{l=1 \ l \neq j}}^{m} \frac{k_l^j \left(\left(d_j \left(s - r_j \right) - c_j - \frac{1}{2d_l + k_l^j} \right) \right)}{2d_l + k_l^j} \right).
$$
\n3.3.3 The Model of the EPA

The central idea of the fund balance is on the basis of balance between the total tax revenue and the total subsidy expenditure. That is,

$$
t'Q_x^{\mu^*} = s'Q_c^{\mu^*}.
$$
\n(3.41)

In order to have a fair basis for comparison, we let the EPA collect the identical total tax revenue under the two different policies. Thus,

$$
t'Q_x^{*} = t^*Q_x^*.
$$
 (3.42)

Substituting (3.5) and (3.37) in (3.42) , we have

$$
nt^2 - \left(na - \sum_{i=1}^n C_{\nu_i}\right)t^2 + \left(\left(na - \sum_{i=1}^n C_{\nu_i}\right)t^* - nt^{*2}\right) = 0. \tag{3.43}
$$

Solving for the roots of (3.43), we obtain the ARF per unit.

$$
t' = \frac{\left(na - \sum_{i=1}^{n} C_{v_i}\right) \pm \sqrt{\left(\sum_{i=1}^{n} C_{v_i} - na + 2nt^*\right)^2}}{2n} \tag{3.44}
$$

Solving (3.45), we obtain

$$
t' = t^* \tag{3.45}
$$

or

$$
t' = a - \frac{\sum_{i=1}^{n} C_{v_i}}{n} - t^*.
$$
\n(3.46)

Substituting (3.26) in (3.45) and (3.46), and let t_1 ['] and t_2 ['] denote these two different ARF rates per unit since it comes with two different roots to (3.43).

$$
t_1' = \frac{1}{n^2} \left((n+1)(nE\tau + ne) - na + \sum_{i=1}^{n} C_{v_i} \right),\tag{3.47}
$$

$$
t_2 = \frac{1}{n^2} \bigg((n+1)na - (n+1)(nEt + ne) - (n+1) \sum_{i=1}^n C_{v_i} \bigg).
$$
 (3.48)

Substituting (3.47) and (3.48) in (3.35) and (3.36) separately and let P_{x_i} ^{*} and q_{x_i} ^{*} denote the *i* th MIS firm's optimal production quantity and market price which respond to t_1 ['], and let P_{x^2} ^{*} and $q_{x_i^2}$ ^{*} denote the *i* th recycling firm's optimal reward money and resulting recycling quantity which respond to t_2 [']. The results are as follows:

$$
P_{x^1}^* = E\tau + e + \frac{\sum_{i=1}^n C_{v_i}}{n},
$$
\n(3.49)

$$
P_{x^2}^{*} = \frac{1}{n(n+1)} \Bigg((n+2)na - (n+1)(nE\tau + ne) - \sum_{i=1}^n C_{v_i} \Bigg), \tag{3.50}
$$

$$
q_{x_i^{1}}^{*} = \frac{1}{(n+1)n^2b} \left(n(n+1)(a - E\tau - e - nC_{v_i}) + (n^2 - 1) \sum_{i=1}^{n} C_{v_i} \right), \tag{3.51}
$$

$$
q_{x_i^2}^{r^*} = \frac{1}{(n+1)n^2b} \left(\frac{-na + (n+1)(nE\tau + ne) - (n+1)n^2C_{v_i} + (n^2 + n + 1)\sum_{i=1}^n C_{v_i}} \right).
$$
 (3.52)

Denoting
$$
D = \frac{\left(-\sum_{i=1}^{n} C_{\nu_i} - ne - nE\tau + na\right)\left(\sum_{i=1}^{n} C_{\nu_i} - na + n(n+1)(e + E\tau)\right)}{n^3b}
$$
 to

simplify our notation and substituting (3.40) , (3.47) , (3.51) in (3.41) , we have

$$
\left(\sum_{\substack{m \ j=1}}^{m} \frac{d_j^2 - d_j \sum_{\substack{l=1 \ l \neq j}}^{m} \frac{k_l^j (d_j + d_l + k_j^l)}{2d_l + k_l^j}}{2d_j - \sum_{\substack{l=1 \ l \neq j}}^{m} \frac{k_l^j (2d_j + k_j^l)}{2d_l + k_l^j}}\right) s^{i2} + \frac{-d_j^2 r_j + d_j c_j + d_j \sum_{\substack{l=1 \ l \neq j}}^{m} \frac{k_l^j (d_j r_j + d_l r_l - c_j + c_l + r_j k_j^l)}{2d_l + k_l^j}}{2d_l + k_l^j} \right) (3.53)
$$
\n
$$
\left(\sum_{j=1}^{m} \frac{-d_j^2 r_j + d_j c_j + d_j \sum_{\substack{l=1 \ l \neq j}}^{m} \frac{k_l^j (2d_j + k_j^l)}{2d_l + k_l^j}}{2d_l + k_l^j}\right) s^i - D = 0.
$$

Denoting
$$
F_j = d_j \sum_{\substack{l=1 \ l \neq j}}^m \frac{k_l^j (d_j + d_l + k_j^l)}{2d_l + k_l^j}, \quad G_j = d_j \sum_{\substack{l=1 \ l \neq j}}^m \frac{k_l^j (d_j r_j + d_l r_l - c_j + c_l + r_j k_j^l)}{2d_l + k_l^j}
$$

to simplify our notation, (3.53) can be simplified to be

$$
\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right) s^{2} + \left(\sum_{j=1}^{m} \frac{-d_j^2 r_j + d_j c_j + G_j}{A}\right) s - D = 0.
$$
\n(3.54)

To ensure the feasibility of (3.54), we use the discriminant of the quadratic polynomial to check this. The testing is as follows:

$$
\left(\sum_{j=1}^{m} \frac{-d_j^2 r_j + d_j c_j + G_j}{A}\right)^2 + 4\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right) D. \tag{3.55}
$$

However, we cannot directly specify whether the value of (3.55), which is influenced by the value of parameters, is positive or not. To avoid a trivial solution in the model, we assume the value of (3.55) is positive. In other words, we do not study the outcome where the value of (3.55) is negative. From this condition follows Assumption 2.

Assumption 2 *The discriminant represented in* (3.55) *is assumed to be positive such that there are rational roots of* (3.54)*; that is, if the value of* (3.55) *is negative, the roots of* (3.54) *are imaginary number.*

Solving the roots of (3.54) and denoting 2 1 $\sum_{j=1}^{m} -d_j^2 r_j + d_j c_j + G_j$ *j* $d_i^2r_i + d_i c_i + G$ *H* $=\sum_{j=1}^{m} \frac{-d_j^2 r_j + d_j c_j + G_j}{A}$ to simplify our

notation, we obtain the subsidy fees per unit.

$$
s' = \frac{-H + \sqrt{H^2 + 4\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right) D}}{2\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)}
$$
(3.56)

Substituting (3.56) in (3.39) and (3.40), we obtain the *j* th recycling firm's optimal reward money and resulting recycling quantity.

$$
P_{w_j}^{\bullet} = \frac{1}{A} \left(1 - \frac{H + \sqrt{H^2 + 4\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)D}}{2\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)} - r_j - c_j - \frac{H}{A} \right) \left(\sum_{j=1}^{m} \frac{K_j^j}{2d_i + k_j^j} \left(\frac{-H + \sqrt{H^2 + 4\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)D}}{d_j r_j + d_i r_i - c_j + c_i} \right) \right) \qquad (3.57)
$$
\n
$$
q_{c_j}^{\bullet} \equiv d_j \left(1 - \frac{d_j}{A} \right) \left(\frac{-H + \sqrt{H^2 + 4\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)D}}{2\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)D} - r_j + \frac{1}{2\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)D} - r_j + \frac{d_j}{2\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)} \right) \qquad (3.58)
$$
\n
$$
\frac{d_j}{A} \left(c_j + \sum_{l=1}^{m} \frac{k_l^j}{2d_l + k_l^j} \left(\frac{-H + \sqrt{H^2 + 4\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)D}}{2\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)} (d_j - d_i) - \frac{1}{2\left(\sum_{j=1}^{m} \frac{d_j^2 - F_j}{A}\right)} \right) \qquad (3.58)
$$

Chapter 4 Case Study and Numerical Analysis

In this chapter, we provide a set of numerical experiments to illustrate the use of the social welfare model and the fund balance model to determine the ARFs and subsidy fees in a competitive closed-loop supply chain in Taiwan and the behavior of the ARFs and subsidy fees with different objectives. Furthermore, we do sensitivity analysis to illustrate possible trends.

4.1 The Case Study and the Numerical Results

We consider the market of laptop computers and assume that there are three MIS firms and two recycling firms in this case study. We estimate the parameters under the reasonable assumption that all the parameters, decision variables, participants' profits, and the externality costs are positive. We follow the case study data presented in (Hong and Ke, 2009) and assume the inverse demand function in the consuming market as $P_x = 33,000 - 0.01 Q_x$, and the recycling quantity functions in the recycling market as $q_{c_1} = 60,000 + 500 P_{w_1}$ 2 1 1 1 w_1 $\sum_{\substack{l=1 \ l \neq 1}}$ $\sum_{\substack{l \neq l}}$ $P_{w_i} - \sum k_l^1 P_i$ $\frac{1}{1}$ $-\sum k_i P_{w_i}$ for the 1st recycling firm

WILLIA,

and $q_{c_2} = 120,000 + 1,000 P_{w_2}$ 2 2 1 2 w_2 $\sum_{\substack{l=1 \ l \neq 2}}$ $\sum_{\substack{l \neq l}}$ $P_{w_2} - \sum k_l^2 P_{\rm g}$ $\frac{1}{1}$ $-\sum k_i^2 P_{w_i}$ for the 2nd recycling firm, where all

currency is in New Taiwan Dollars (NTD). We assume that the production cost per unit of new laptop computers for the three MIS firms is 25,000/26,000/27,000 NTD. It is not straightforward to estimate the unit indirect pollution cost, *e* , incurred in producing new products and the unit pollution cost of uncollected e-scrap products, *E* . According to Li (2005), production costs have increased by around 5-10% due

to the launch of WEEE and RoHS. In this case study, we estimate the unit indirect pollution cost, *e* , as the average increase (7.5%) in the three MIS firms' average production costs due to WEEE and RoHS ($e = 1,950$). In Wen (2005b), the total cost of recycling one unit of laptop computers is approximately estimated as 135 NTD, which is assumed to be the unit pollution cost of uncollected e-scrap products, *E* . The ballpark figure of the value of recovered components of laptop computers is approximately estimated as 83 NTD (Wen, 2005b). Therefore, the net cost for recycling one unit of laptop computers is approximated to 50 NTD. The rate of current generation of new products that are expected to be returned to the reverse channel after usage is estimated as 0.97% (Wen, 2005b). We note that the return rate in this case study is relatively lower than our initial conjecture. This is probably because most consumers retain obsolete laptop computers because of their relatively high price and small volume compared to desktop computers.

Besides, we set the decrease in the recycling quantity in the *j* th recycling firm caused by a unit of increase in the reward money paid by the *l* th recycling firm, k_1^2 and k_2^1 in this case study, as 10 or 20. In this case study, we call the scenario where the value of k_1^2 and k_2^1 is 10 the 1st scenario, and the scenario where the value of k_1^2 and k_2^1 is 20 the 2nd scenario. The 1st scenario represents a small influence of a recycling firm's reward money on another recycling firm's recycling quantity, and the 2nd scenario represents a large influence of a recycling firm's reward money on another recycling firm's recycling quantity. Due to a fair basis for comparison, we refer to t_1 ', which is the same as the ARF in the social welfare model, as the ARF in the fund balance model in this case study. The estimated data are summarized in Table 1.

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Table 1 Parameters in the Numerical Study

Based on the estimated data in Table 1, the subsidy fee in the social welfare model and the fund balance model is 193 NTD and 392 NTD respectively in the 1st scenario, and 383 NTD and 392 NTD respectively in the 2nd scenario. Because we have assumed that there are identical tax revenues under the two different models due to a fair basis for comparison, the computed ARF for both models in both scenarios is 268 NTD. In this case study, the value of social welfare in the social welfare model obviously outperforms the welfare value in the fund balance model in both scenarios. Our results show that social welfare improves by approximately 1.15% in the 1st scenario and 0.08% in the 2nd scenario if the EPA chooses a welfare maximization model instead of the fund balance model in the laptop computer market. Furthermore, we use sensitivity analysis to investigate the impact of these parameters on the value of social welfare by the estimated data.

4.2 Sensitivity Analysis

We first study how the characteristics of the consuming market affect the value of social welfare, where the consuming market characteristics can be interpreted as the parameters, *a* and *b* , in the inverse demand function. The results are given in Figure 3. We summarize the major observations as follows:

- (i) All other parameters remaining the same, for the value of social welfare, the social welfare model outperforms the fund balance model.
- (ii) An increase in *a* results in an increase in the value of social welfare for both models in both scenarios. This indicates a positive relation between social welfare and a . However, both models in both scenarios show that social welfare decreases as the value of *b* increases, implying a negative relation.
- (iii) The difference in the value of social welfare between the models decreases in both scenarios as the value of *a* increases, and the difference in the value of social welfare between the models is small in both scenarios as the value of *b* increases. This shows that the EPA may pay more attention to the best way to determine the associated fees when the value of a is at a low level, since

the fund balance model may give a poor performance at the low level of the value of *a* , especially in the 1st scenario; that is, the decrease in the recycling quantity in the 2nd/1st recycling firm caused by a unit of increase in the reward money paid by the 1st/2nd recycling firm is small.

Figure 3 Impact of *a* and *b* on the value of social welfare

We next study the impact of the unit indirect pollution cost incurred in producing new products (e) and the unit pollution cost of uncollected e-scrap products (E) on the value of social welfare. The results are given in Figure 4. We summarize the major observations as follows:

- (i) All other parameters remaining the same, for value of social welfare, the social welfare model outperforms the fund balance model.
- (ii) An increase in *e* results in a decrease in the value of social welfare for both models in both scenarios. This indicates a negative relation between social

welfare and *e*. However, the social welfare model in both scenarios show that social welfare increases as the value of *E* increases, implying a positive relation, while the fund balance model in both scenarios show that social welfare decreases as the value of *E* increases, implying a negative relation. It is surprising that social welfare increases as the unit pollution cost of uncollected e-scrap products (E) increases in the social welfare model. From the numerical results, we observe that EPA's profits increases as the value of *E* increases in the social welfare model in both scenarios. A possible explanation is that the EPA raises the ARF to encourage the MIS firms to produce environmental friendly products and an increase in the ARF benefits the EPA's profits in the social welfare model.

- (iii) The difference in the value of social welfare between the models increases as the value of *E* or *e* increases. This shows that the EPA may pay more attention to system objectives when the value of E or e is at a relatively high level, since the fund balance model may perform worse at a relatively high level of the value of E or e .
- (iv) For both E and e , the value of social welfare in the 2nd scenario is smaller than in the 1th scenario in the social welfare model while all other parameters are remaining the same. From the numerical results, we observe that the reward money of each recycling firm in the 2nd scenario is higher than in the 1th scenario. It is reasonable to imagine that the EPA would raise the subsidy fee when recycling firms raise their reward money and an increase in the subsidy fees decreases the EPA's profits.

Figure 4 Impact of *e* and *E* on the value of social welfare

We next study the impact of return rate, τ , on the value of social welfare. The results are given in Figure 5. We summarize the major observations as follows:

- (i) All other parameters remaining the same, an increase in the value of τ implies an increase in the recycling quantity in the recycling stream. In this situation, the EPA can choose to raise the ARF to restrain the consumption of new products in the social welfare model and raise the ARF to balance the increase in the expenditure of subsidy fee in the fund balance model. As a result, the value of the social welfare decreases for both models in both scenarios.
- (ii) In the social welfare model in both scenarios, there is a turning point in the curve of the value of social welfare. From the numerical results, we observe that the number of uncollected e-scrap products is zero when the return rate is

less than a certain number. That is, the externality cost of uncollected e-scrap products is zero when the return rate is less than a certain number. In this numerical example, the certain number is 0.5 in the 1st scenario and 0.7 in the 2nd scenario.

(iii) All other parameters remaining the same, the value of social welfare in the 2nd scenario is smaller than in the 1st scenario in the social welfare model. From the numerical results, we observe that the reward money of each recycling firm in the 2nd scenario is higher than in the 1st scenario. It is reasonable to imagine that when recycling firms raise their reward money, the EPA would raise the subsidy fee, and then an increase in the subsidy fees decreases the

Figure 5 Impact of τ on the value of social welfare

4.3 Analysis of the Recycling Market

In this section, we study how the market intercept parameter and slope parameter of the recycling market affect the reward money, recycling quantity, subsidy fee, and the value of social welfare. As mentioned in Section 3.2.2, the situations in different areas cannot be simplified as one single market, so we let c_j and d_j denote the *j* th recycling firm's market intercept parameter and slope parameter respectively, $j = 1, 2$, in this case study.

We first study the intercept parameters, c_1 and c_2 . The parameter, c_j , denotes the *j* th recycling firm's basic recycling quantity when the reward money of all recycling firms is zero. Hence, we denote the economic meaning of c_i be the size of the market area where the *j* th recycling firm exists. In this case study, we assume that the 1st recycling firm exists in a small market area and the 2nd recycling firm exists in a big market area. Therefore, we use the ratio of c_2 to c_1 , c_2/c_1 , as the ratio of the size of the market areas where the 1st and 2nd recycling firms exist. We assume that the sum of c_1 and c_2 is fixed due to a fair basis for comparing between different scenarios, where all the parameters are remaining the same but c_2/c_1 are different. Because the 1st scenario and 2nd scenario behave in a similar manner, we only set k_1^2 and k_2^1 as 10 in this case study; that is, there is only the 1st scenario considered in this case study for simplicity. We use parts of the data in Table 1. The results are given in Figure 6. We summarize the major observations as follows:

- (i) For the 1st recycling firm in both models, the reward money increases but the recycling quantity decreases as c_2/c_1 increases. However, for the 2nd recycling firm in both models, the reward money decreases but the recycling quantity increases as c_2/c_1 increases. That is, although the 2nd recycling firm decreases its reward money, consumers are willing to bring the obsolete products to the 2nd recycling firm.
- (ii) In the fund balance model, there is no obvious change in the subsidy fees and

the value of social welfare as c_2/c_1 increases. However, in the social welfare model, the subsidy fees decrease as c_2/c_1 increases. Besides, a decrease in subsidy fees benefits the EPA's profits, and then an increase in EPA's profits benefits the social welfare.

The ratio of competition between the 1st and 2nd recycling firms (C_2 / C_1)

Figure 6 Impact of c_2/c_1 on reward money, recycling quantity, subsidy fees, and the value of social welfare

We next study the slope parameters, d_1 and d_2 . The parameter, d_j , denotes the increase in the number of recycling quantity of the *j* th recycling firm when per unit of the *j* th recycling firm's reward money increases. Hence, we denote the economic meaning of d_i be the level of influence on the recycling market. In this case study, we assume that the 2nd recycling firm has more powerful influence on the recycling market than the 1st recycling firm. Therefore, we use the ratio of d_2 to d_1 , d_2/d_1 , as the ratio of market influence between the 1st and 2nd recycling firms. We assume that the sum of d_1 and d_2 is fixed due to a fair basis for comparing between different scenarios, where all the parameters are remaining the same but d_2 / d_1 are different. Because the 1st scenario and 2nd scenario behave in a similar manner, we only set k_1^2 and k_2^1 as 10 in this case study; that is, there is only the 1st scenario considered in this case study for simplicity. We use parts of the data in Table 1. The results are given in Figure 7. We summarize the major observations as follows:

- (i) For the 1st recycling firm in both models, the reward money and recycling quantity decrease as d_2/d_1 increases. However, for the 2nd recycling firm in both models, the reward money and recycling quantity increase as d_2/d_1 increases. It is reasonable to illustrate that the reward money and recycling quantity of a recycling firm increase when the market influence of the recycling firm increases.
- (ii) In the fund balance model, there is no obvious change in the subsidy fees and the value of social welfare as d_2/d_1 increases. However, in the social welfare model, the subsidy fees increase as d_2 / d_1 increases. Moreover, an increase in subsidy fees decreases the EPA's profits, and then a decrease in EPA's profits decreases the social welfare.

 \rightarrow The social welfare model-recycling firm 1 \rightarrow The social welfare model-recycling firm 2 \rightarrow The fund balance model-recycling firm 1 \rightarrow The fund balance model-recycling firm 2

Chapter 5 Recycling Licenses

The general sequence of setting up a recycling firm is that a recycling firm has to be granted a legal recycling license by the government, make blueprint and detailed constructing plans, gradually establish business network and accumulate government support (Wang, 2008). For example, the Taiwan recycling firms are required to be granted recycling licenses and then subsidized by the EPA (RFMB, 2009d). Hence, the EPA has the political power to decide how many recycling firms exist in the recycling market. From the EPA's standpoint, it may consider the total social welfare when it makes policies. The number of recycling firms in the recycling market may affect the value of social welfare, so it is important for the EPA to determine the number of recycling licenses in the recycling market. In this chapter, we study how the EPA determines the optimal number of recycling licenses in the recycling market to maximize the social welfare, and study the impact of the number of recycling licenses on the value of social welfare, total recycling quantity, reward money, and subsidy fees for the social welfare model and the fund balance model.

5.1 The Model of Homogeneous Recycling Firms

In this section, we study how the EPA determines the optimal number of recycling licenses in the recycling market to maximize the social welfare. From Chapter 3, we know that recycling firms usually collect e-scrap products in distinct market segments; that is, when there are *m* areas in the recycling market, there are *m* recycling firms in the recycling market. In other words, the EPA has the political power to determine geographically exclusive areas in the recycling market; that is, the EPA has to divide the recycling market into different areas, where only one recycling firm

exists in one area and it is responsible for recycling.

In Chapter 3, we assume that recycling firms are heterogeneous. For simplicity, in this chapter, we assume that recycling firms are homogeneous. For the recycling firms, the net cost for recycling one unit of e-scrap products, r , is the same, $r > 0$. A decrease in the recycling quantity in a recycling firm caused by a unit of increase in the reward money paid by another recycling firm, k is the same between any two recycling firms. As mentioned in Chapter 3, we know that a legal recycling firm exists in one area and proceeds with recycling, and under the assumption that recycling firms are homogeneous, we assume that the situations in different areas are the same. We let *c* and *d* denote the intercept and slope parameters of the recycling quantity function of each recycling firm in the recycling market. Furthermore, we use the parameters and assumptions mentioned above to solve for the policies of each participant under the social welfare model and fund balance model.

5.1.1 The Social Welfare Model

In the social welfare model, the EPA aims to maximize the total social welfare when it makes policies. Then the MIS firms and recycling firms aim to maximize their profits according to the level of ARFs and subsidy fees announced by the EPA.

In this chapter, we study the impact of the number of recycling licenses in the recycling market, so the production quantity of the MIS firms and the level of ARFs, *t*, are the same as the analytical solutions in Section 3.2.3. Let P_w and q_c denote the reward money and recycling quantity of each recycling firm respectively. Given that the EPA issues *m* recycling licenses in the recycling market; that is, the EPA divides the recycling market into *m* areas. According to (3.9) and the assumption that recycling firms are homogeneous, the recycling quantity function is listed as follows:

$$
q_c = c + (d - mk + k)P_w.
$$
\n
$$
(5.1)
$$

According to (3.32), (3.33), and the assumption that recycling firms are homogeneous, the reward money and recycling quantity of each recycling firm are written as follows:

$$
P_w^* = \frac{1}{d}(E - r)(d - mk + k),
$$
\n(5.2)

$$
q_c^* = c + \frac{1}{d}(E - r)(d - mk + k)^2.
$$
 (5.3)

Adding up recycling firms' recycling quantity together, we obtain the total recycling quantity, Q_c , as follows: ≣TEISI

$$
Q_c = m \left(c + \frac{1}{d} (E - r) (d - mk + k)^2 \right).
$$
 (5.4)

The EPA maximizes the total social welfare while determining the level of the ARFs and subsidy fees. Under the assumption that recycling firms are homogeneous, the social welfare defined in Section 3.2.3 can be written as follows:

$$
\sum_{i=1}^{n} (P_x - C_{\nu_i} - t) q_{x_i} + m(s - P_w - r) q_c + \frac{1}{2} b Q_x^2 +
$$

\n
$$
m (P_w q_c - \frac{1}{2} d P_w^2) + (t Q_x - m(s q_c)) - E(\tau Q_x - m q_c) - e Q_x.
$$
\n(5.5)

According to (3.27) and the assumption that recycling firms are homogeneous, the level of the subsidy fees, *s* , is written as follows:

$$
s^* = \frac{(E-r)(2d-mk+k)(d-mk+k)}{d^2} + \frac{dr+c}{d}.
$$
 (5.6)

5.1.2 The Fund Balance Model

In the fund balance model, which is the current practice model, the EPA determines the level of the ARFs and subsidy fees on the basis of balance between the total tax revenue and the total subsidy expenditure. Then the MIS firms and recycling firms aim to maximize their profits according to the level of fees announced by the EPA.

As mentioned in Section 5.1.1, the production quantity of the MIS firms and the level of ARFs, *t* , are the same as the analytical solutions in Section 3.3.3. The parameters, r , k , c , d , in this section are the same as the parameters in Section 5.1.1. Let P_w and q_c denote the reward money and recycling quantity of each recycling firm respectively. Given that the EPA issues m' recycling licenses in the recycling market; that is, the EPA divides the recycling market into m' areas. According to (3.9) and the assumption that recycling firms are homogeneous, the recycling quantity function is listed as follows:

$$
q_c' = c + (d - m'k + k)P_w'. \tag{5.7}
$$

According to (3.57), (3.58), and the assumption that recycling firms are homogeneous, the reward money and recycling quantity of each recycling firm are written as follows:

$$
P_w^* = \frac{\left(\frac{dm'(c - (d - m'k + k)r)}{(2d - m'k + k)} + \frac{m'^2 d^2 (c - (d - m'k + k)r)^2}{(2d - m'k + k)} + \frac{m'^2 d^2 (c - (d - m'k + k)r)^2}{(2d - m'k + k)^2}\right)}{2m'(d - m'k + k)}
$$
(5.8)

$$
\frac{dr + c}{(2d - m'k + k)},
$$

$$
q_c^* = c + -\frac{d(c - (d - m'k + k)r)}{2(2d - m'k + k)} +
$$

\n
$$
\frac{1}{2m'} \sqrt{\frac{4Dm'd(d - m'k + k)}{(2d - m'k + k)} + \frac{m'^2 d^2(c - (d - m'k + k)r)^2}{(2d - m'k + k)^2}} -
$$
\n
$$
\frac{(d - m'k + k)}{(2d - m'k + k)} (dr + c).
$$
\n(5.9)

Adding up recycling firms' recycling quantity together, we obtain the total recycling quantity, Q_c ['], as follows:

$$
Q_c^* = m'c + \frac{dm'(c - (d - m'k + k)r)}{2(2d - m'k + k)} +
$$

\n
$$
\frac{1}{2}\sqrt{\frac{4Dm'd(d - m'k + k)}{(2d - m'k + k)} + \frac{m'^2d^2(c - (d - m'k + k)r)^2}{(2d - m'k + k)^2}} -
$$

\n
$$
\frac{m'(d - m'k + k)}{(2d - m'k + k)}(dr + c).
$$
\n(5.10)

The EPA determines the level of the ARFs and subsidy fees on the basis of balance between the total tax revenue and the total subsidy expenditure. In order to have a fair basis for comparison, we study the impact of the number of recycling licenses on the value of social welfare, total recycling quantity, reward money, and subsidy fees on the basis of fund balance in this section. Under the assumption that recycling firms are homogeneous, the social welfare defined in Section 3.2.3 is as follows:

$$
\sum_{i=1}^{n} (P_x - C_{\nu_i} - t) q_{x_i} + m'(s' - P_{\nu} - r) q_c + \frac{1}{2} b Q_x^2 +
$$
\n
$$
m'\Big(P_{\nu} \Big| q_c \Big| - \frac{1}{2} d P_{\nu} \Big|^2 + \Big(t Q_x - m'(s' q_c) \Big) - E\Big(\tau Q_x - m' q_c \Big) - e Q_x.
$$
\n(5.11)

According to (3.56) and the assumption that recycling firms are homogeneous, the level of the subsidy fees, *s*' , is written as follows:

$$
\begin{bmatrix}\n-dm'(c - (d - m'k + k)r) + \\
(2d - m'k + k) \\
\frac{4Dm'd(d - m'k + k)}{(2d - m'k + k)} + \frac{m'^2 d^2(c - (d - m'k + k)r)^2}{(2d - m'k + k)^2} \\
x^* &= \frac{\sqrt{4Dm'd(d - m'k + k)} + \frac{m'^2 d^2(c - (d - m'k + k)r)^2}{(2d - m'k + k)^2}}{2m'd(d - m'k + k)}.\n\end{bmatrix}
$$
\n(5.12)

5.1.3 The Optimal Number of Recycling Licenses

In this section, we study how the EPA determines the optimal number of recycling licenses in the recycling market to maximize the social welfare. In the social welfare model, the EPA aims to maximize the total social welfare. All other parameters remaining the same, the value of (5.5) may increase or decrease as the value of *m* increases. Therefore, a value, m^* , which satisfies the maximum of (5.5), is the optimal number of recycling licenses in the recycling market in the social welfare model. However, in the fund balance model, the EPA aims to establish the level of the ARFs and subsidy fees on the basis of balance between the total tax revenue and the total subsidy expenditure. For a fair basis for comparison, we let the value of social welfare be the performance measure in the fund balance model. All other parameters remaining the same, the value of (5.11) may increase or decrease as the value of m' increases. Therefore, a value, m^* , which satisfies the maximum of (5.11), is the optimal number of recycling licenses in the recycling market in the fund balance model.

 Furthermore, we utilize a set of numerical experiments to study how the EPA determines the optimal number of recycling licenses in the recycling market to maximize the social welfare and illustrate the impact of the number of recycling licenses on the value of social welfare, total recycling quantity, reward money, and subsidy fees for the social welfare model and the fund balance model.

5.2 Case Study

We let *m* denote the number of recycling firms in the recycling market in this section, and use parts of the data in Table 1. We set *k* as 10, *c* as 90,000, and *d* as 750, where *c* is the average value of c_1 and c_2 , and *d* is the average value of d_1 and d_2 . On the other hand, c is viewed as the potential recycling quantity in a recycling area while the reward money paid by the recycling firm is zero; moreover, the number of market areas in the recycling market is decided by the EPA. Therefore, we reasonably assume that the potential recycling quantity in a recycling area is *c* while the EPA does not divide the recycling market, and the potential recycling quantity in a recycling area is $c/2$ while the EPA divides the recycling market into two areas, and so on. In this case study, *c* is 90,000 while *m* is 1, and *c* is 45,000 while *m* is 2, and so on. Based on the estimated data mentioned above, we obtain the possible trends, which are given in Figure 8. We summarize the major observations as follows:

- (i) All other parameters remaining the same, an increase in *m* results in a decrease in the reward money in the social welfare model, and an increase in *m* first results in a decrease and then leads to an increase in the reward money in the fund balance model.
- (ii) All other parameters remaining the same, an increase in *m* results in a decrease in the level of the subsidy fees in the social welfare model, and an increase in *m* first results in a decrease and then leads to an increase in the

level of the subsidy fees in the fund balance model.

- (iii) All other parameters remaining the same, an increase in *m* first results in an increase and then leads to a decrease in total recycling quantity in both models.
- (iv) For the social welfare model, an increase in *m* first results in a decrease and then leads to an increase in the value of social welfare. From the numerical results, we observe that total recycling quantity first increases and then decreases as *m* increases, so the EPA's profits first increases and then decreases as *m* increases, and then an increase/decrease in the EPA's profits benefits/decreases the social welfare. This shows that the EPA may approve as many as applications for recycling licenses in the social welfare model. However, the fund balance model shows that an increase in *m* results in a decrease in the value of social welfare. From the numerical results, we observe that the profits of recycling firms decrease as *m* increases, and a decrease in the profits of recycling firms decreases the social welfare. This shows that the EPA may not approve many applications for recycling licenses to obtain better value of social welfare in the fund balance model.

The number of recycling firms in the recycling market (*m*)

Figure 8 Impact of *m* on subsidy fees, reward money, total recycling quantity, and the value of social welfare

Chapter 6 Conclusion and Future Research

Several environmental regulations are announced due to high demands in many raw material markets and growing concerns about the environmental impacts of disposal. In the light of management and recycling of e-scrap products, the EPA imposes taxes on MIS firms to restrain excessive production of new products and produce environmental friendly products, and compensate recycling firms for the costs along with recycling and processing the e-scrap products to encourage recycling programs.

This thesis presents the Stackelberg-type model, which is the social welfare model. The EPA is a leader to determine the level of the ARFs and subsidy fees to maximize the social welfare. The MIS firms and recycling firms are followers, who choose the optimal production quantity of new products and the optimal reward money after observing the level of fees announced by the EPA. We assume that there is a competitive market in closed-loop supply chains; that is, there are many MIS firms and recycling firms in the consuming market and recycling market respectively. Currently, the EPA determines the level of the ARFs and subsidy fees on the basis of fund balance between the total tax revenue and total subsidy expenditure. For comparative purposes, we also develop the fund balance model where the total tax revenue is equal to the total subsidy expenditure. Then we examine the numerical study to illustrate the use of the social welfare model and the fund balance model. Besides, we do sensitivity analysis to illustrate possible trends.

Furthermore, we assume that recycling firms are homogeneous. We study how the EPA determine the optimal number of recycling licenses in the recycling market to maximize the social welfare and study the impact of the number of recycling licenses on the value of social welfare, total recycling quantity, reward money, and subsidy fees.

We summarize our results as follows:

- (i) The proposed social welfare model outperforms the fund balance model considering the value of social welfare in a competitive system.
- (ii) The difference in the value of social welfare between the social welfare model and the fund balance model is big when one of the following conditions holds: the market price is high, the environmental externality cost is high, or the influence between the recycling firms is big. That is, the improvement in the value of social welfare is great when one of the three conditions mentioned above holds.
- (iii) The level of ARFs is the same in the social welfare model and the fund balance model. The level of subsidy fees is lower in the social welfare model than the level of subsidy fees in the fund balance model.
- (iv) The optimal production quantity of the MIS firms is the same in the social welfare model and the fund balance model. The reward money of the recycling firms is lower in the social welfare model than the reward money of the recycling firms in the fund balance model.
- (v) Considering the value of social welfare, the EPA may approve as many as applications for recycling licenses in the social welfare model, but may not approve many applications for recycling licenses in the fund balance model.

In this thesis, we assume that recycling firms exist in distinct market areas and there is only one recycling firm in a market area. It would be interesting to investigate another situation where all recycling firms exist in a recycling market. On the other hand, we only assume that the intercept parameter decreases as the number of recycling licenses increases when we study the number of recycling licenses. However, other parameters, such as the influence between the recycling firms, may increase or decrease as the number of recycling licenses increases. It would be more realistic to consider the change of other parameters as the number of recycling licenses increases.

In this thesis, we describe the demand function and the recycling quantity function as linear functions. It would be interesting to develop a model where the demand function and recycling quantity function of different products are different. In addition, there is only one period in this thesis; moreover, we do not consider the issues of inventory and constraints on productivity. It would be more realistic to consider the issues of inventory or constraints on productivity, and then develop a suitable model for studying the recycling system over many periods.

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